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# Edge Effects on Stand Structure and Regeneration in a Subalpine Coniferous Forest on Mt. Fuji, Japan, 30 Years after Road Construction

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## Abstract

To evaluate edge effects in subalpine coniferous forest, I quantified stand structure characteristics in a plot that extended into the forest from a road that had been constructed approximately 30 yr previously on Mt. Fuji, in central Japan. The basal area of standing dead trees and the density of saplings in the plot were greater near the road. However, cluster analysis indicated that stand structure near the road edge was similar to that of gaps in the interior of the stand. Since edge effects may result from interactions among several variables, analyzing data sets that include several variables rather than a single variable is a promising approach for evaluating edge effects. In a multiple regression analysis, sapling densities of *Tsuga diversifolia* and *Abies veitchii* were strongly affected by the basal area of standing dead trees that had apparently died as a result of the construction. *A. veitchii* is more shade intolerant than the other two main species in the study plot, but it responds quickly to increased light availability. Although the plot is currently composed of *T. diversifolia* and *A. mariesii*, successful regeneration of *A. veitchii* may eventually change the species composition of this plot.

## Introduction

Road construction changes the spatial patterns of a forested landscape. The process, defined as dissection (Forman, 1995), creates an edge and severely affects the biotic and abiotic environments of the surrounding areas (e.g., Schonewald-Cox and Buechner, 1992; Noss and Csuti, 1994; Young, 1994). These edge effects, which are caused not only by road construction but also by other human impacts that transform a landscape, such as clear-cutting, have been widely reported (e.g., Harris, 1984; Brothers and Spingarn, 1992; Fraver, 1994; Young and Mitchell, 1994). Evaluations of edge effects have varied, depending on the species and scale context, from their being considered a traditionally positive effect, especially for wildlife (see Yahner, 1988), to being considered a negative effect on biodiversity (e.g., Noss, 1983; Harris, 1988; Chen et al., 1992). Most research on edge effects has focused on measuring the extent of their influence (e.g., Williams-Linere, 1990; Matlack, 1993).

It is important for management and restoration purposes to measure the “edge depth,” i.e., the distance from the edge to the forest interior, to maximize the amount of interior habitat (Laurance and Yensen, 1991; Sisk and Margules, 1993). However, because edge effects may be caused by interactions among several variables, including biotic and abiotic parameters, edge depth is difficult to define and to measure (Murcia, 1995; Cadenasso et al., 1997; Esseen and Renhorn, 1998). In addition, because edge effects do not have clear boundaries, the influence of an edge follows a gradient (Fraver, 1994). Since species respond to edges in different ways (Palik and Murphy, 1990; Matlack, 1994), Murcia (1995) suggested that any measurement of edge effects should consider which spatial and temporal scales are most relevant for each species. Some researchers (Murcia, 1995; Cadenasso et al., 1997; Esseen and Renhorn, 1998; McCollin, 1998) have pointed out that in a large proportion of existing studies, the definition of an edge is not based on functional properties.

A plant community consists of a mosaic of patches at different stages of development (Watt, 1947). Mosaic structures, consisting of patches that vary in size, age structure, and developmental stage, have

been identified in many forest types (e.g., Nakashizuka, 1983; Veblen, 1986; Liu and Hytteborn, 1991; Lertzman, 1992; Yamamoto, 1992a; Hiura et al., 1996). The development and regeneration of forest starts from a disturbance (e.g., Yamamoto, 1992b; Runkle, 1998). To show how patches develop, various phases of forest development must be compared. Therefore, any comparison of stand structure along the edge and in the interior should consider the mosaic structure of the interior. Previous research on edge effects in forests has paid little attention to this point. To demonstrate the stand structure of an edge, it must be compared to the interior and to different phases of stand development within the interior. In addition, it is important to predict the dynamics, development, and restoration of the stand structure of an edge; to date, however, few studies have evaluated stand development along edges.

Mt. Fuji is the highest mountain in Japan (3776 m a.s.l.). Subalpine coniferous forests dominated by *Tsuga diversifolia*, *Abies mariesii*, and *A. veitchii* cover the north slope from 1500 to 2400 m (Franklin et al., 1979; Ohsawa, 1984). Stand structure and regeneration in the forest are affected primarily by gap formation, which creates a mosaic of patches (Nakamura, 1985). Kanzaki (1984) showed that the canopy layer of a *T. diversifolia* forest in central Japan comprised a mosaic of patches arising from gaps that had formed at different times.

In 1964, a road for sightseeing was constructed through the primary coniferous forest in the subalpine zone of Mt. Fuji. The subsequent death of canopy trees in the edges alongside the road was attributed to road construction (Maeda et al., 1998). Approximately 30 yr after the road was constructed, Maeda et al. (1998) showed that conifers had successfully regenerated, especially along the edges. However, to better understand the process of forest development that has shaped the current stand structure, it is necessary to quantify the regeneration and recovery of the stand.

Using stand structure variables, I quantified the edge effects in this forest. Specifically, my objectives were to determine (1) the phase of forest development near the edge and (2) whether species differed in their regeneration responses following the death of canopy trees.

TABLE 1  
Structure of live and standing dead trees of the stand

|                                  | Density              |       |       |       |            |       |                 |       |           |        |            |            |
|----------------------------------|----------------------|-------|-------|-------|------------|-------|-----------------|-------|-----------|--------|------------|------------|
|                                  | Basal area           |       | Total |       | Saplings** |       | Larger trees*** |       | DBH* (cm) |        |            |            |
|                                  | (m <sup>2</sup> /ha) | (%)   | (/ha) | (%)   | (/ha)      | (%)   | (/ha)           | (%)   | (Mean)    | (Max.) | (S.D.****) | (Skewness) |
| Live trees                       |                      |       |       |       |            |       |                 |       |           |        |            |            |
| <i>Tsuga diversifolia</i>        | 34.9                 | 53.2  | 1179  | 4.4   | 366        | 35.6  | 203             | 55.9  | 13.7      | 54.6   | 13.7       | 0.931      |
| <i>Abies mariesii</i>            | 26.0                 | 39.7  | 1106  | 9.4   | 150        | 14.6  | 139             | 8.2   | 13.5      | 48.5   | 12.2       | 0.955      |
| <i>Abies veitchii</i>            | 4.3                  | 6.5   | 761   | 22.2  | 260        | 25.3  | 21              | 5.9   | 5.4       | 47.7   | 6.5        | 3.661      |
| <i>Rhododendron brachycarpum</i> | 0.4                  | 0.6   | 471   | 13.8  | 251        | 24.4  | —               | —     | 3.0       | 8.6    | 1.3        | 0.750      |
| <i>Betula ermanii</i>            | 0.0                  | 0.0   | 3     | 0.1   | 1          | 0.1   | —               | —     | 4.3       | 6.3    | 2.1        | —          |
| <i>Euonymus macropterus</i>      | 0.0                  | 0.0   | 3     | 0.1   | —          | —     | —               | —     | 5.7       | 7.2    | 1.5        | —          |
| Total                            | 65.6                 | 100.0 | 3423  | 100.0 | 1028       | 00.0  | 363             | 100.0 | 10.3      | 54.6   | 11.7       | 1.507      |
| Standing dead trees              |                      |       |       |       |            |       |                 |       |           |        |            |            |
| <i>Tsuga diversifolia</i>        | 3.3                  | 27.3  | 113   | 5.9   | 7          | 15.2  | 9               | 22.2  | 16.3      | 48.1   | 9.9        | 0.426      |
| <i>Abies mariesii</i>            | 2.4                  | 20.4  | 64    | 14.8  | —          | —     | 7               | 18.5  | 20.0      | 38.4   | 8.8        | 0.126      |
| <i>Abies veitchii</i>            | 0.3                  | 1.8   | 6     | 1.3   | —          | —     | —               | —     | 21.4      | 27.5   | 5.2        | −0.530     |
| <i>Abies</i> spp.                | 4.3                  | 36.1  | 124   | 8.5   | 1          | 3.0   | 21              | 55.6  | 17.6      | 43.8   | 11.3       | 0.430      |
| Conifer spp.                     | 1.7                  | 14.0  | 63    | 14.4  | —          | —     | 1               | 3.7   | 17.1      | 31.4   | 6.5        | 0.309      |
| <i>Rhododendron brachycarpum</i> | 0.0                  | 0.4   | 66    | 15.1  | 39         | 81.8  | —               | —     | 2.7       | 1.0    | 0.9        | 0.302      |
| Total                            | 12.0                 | 100.0 | 436   | 100.0 | 47         | 100.0 | 38              | 100.0 | 15.4      | 48.1   | 10.5       | 0.466      |

\* DBH: Diameter at breast height, \*\* Saplings: DBH <3 cm, \*\*\* Larger trees: DBH ≥30 cm, \*\*\*\* SD: Standard deviation.

## Study Site and Methods

The study plot was located 2100 m a.s.l. on the north slope of Mt. Fuji (35° N, 138°E). Nakamura and Obata (1982) found that the mean monthly temperature on Mt. Fuji (at 1900 m) ranged from 16.4°C in August to −7.6°C in January, the mean annual temperature was 4.3°C, and the mean annual precipitation measured 2300 mm. These values indicate that the study site is located in the subalpine zone. However, the maximum snow depth was approximately 1 m, suggesting that the study site is not in the region of Japan that experiences very heavy snowfall. Before the road was constructed in 1964, the forest was undisturbed. Unfortunately, stand structure and species composition along the road were not recorded prior to construction.

The road is 29.5 km long and climbs from 857 to 2305 m a.s.l.; about 15 km is in subalpine forest. Many canopy trees along the road have died, apparently as a result of the road construction (Maeda et al., 1998). Although the reasons cannot be definitively proven, exposure to strong winds can cause die-backs following road construction. Because standing dead trees along the road were considered unsightly, they were cut down; thus, most of the forest along the road is no longer in a natural condition, with standing dead trees. Standing dead trees were left uncut in only a small part of the forest along the road, to serve as a control site.

I established my study plot in this control site in 1999. The plot was 50 × 140 m (0.7 ha), with the long side of the plot perpendicular to the road. A zone about 10 m wide along the roadside was not included in the plot because the road construction had destroyed the slope. The plot was divided into a 10 × 10-m grid (70 quadrats in total). Since the mean canopy gap size in primary *Abies-Tsuga* forests located 150 km north of this study site was 38.3–84.3 m<sup>2</sup> (Yamamoto, 1995; Narukawa and Yamamoto, 2001), this grid size was considered suitable to analyze the developmental phases of stand structure. Both live trees and standing dead trees >2 m tall were tagged and identified within each quadrat. Fallen dead trees were not measured. Several standing dead trees could not be identified because of bark decomposition. I measured the diameter at breast height (DBH) using a steel tape measure, or diameter vernier calipers for smaller trees. Saplings were defined as trees with a DBH of <3 cm and “large trees” as trees with a DBH ≥30 cm. For *Rhododendron brachycarpum*, a shrub with many stems, only the largest stem of an individual was tagged and

measured. Several studies have indicated that *Rhododendron* spp. impede tree regeneration (e.g., Baker and Van Lear, 1998). However, I found no significant correlations between the individual density of *R. brachycarpum* and conifer sapling density (*T. diversifolia*,  $r = -0.044$ ,  $P = 0.801$ ; *A. mariesii*,  $r = -0.119$ ,  $P = 0.548$ ; and *A. veitchii*,  $r = -0.176$ ,  $P = 0.344$ ), so any effects of *R. brachycarpum* on regeneration were omitted from the analysis. A bryophyte, *Hylocomium splendens*, dominated the forest floor.

Statistical analyses (e.g., cluster analysis and simple correlations among stand attributes) were confined to the three dominant conifers (*T. diversifolia*, *A. mariesii*, and *A. veitchii*). Analyses were performed separately for each quadrat. Cluster analysis was used to examine the phase of forest development and regeneration near the edge, using a data set that included the basal areas of both live and standing dead trees and the densities of both large live trees (DBH ≥30 cm) and saplings of the three dominant species. The dendrogram using the average linkage between groups was separated into four groups at 15% of the squared Euclidean distance (similarity). The four groups were randomly assigned the letters A, B, C, and D. To show whether there were species differences in the regeneration response following the death of canopy trees, I used a simple correlation for each dominant species. Also, multiple regression analysis was performed with sapling density as the dependent variable and stand structure parameters (the basal area of both live trees and standing dead trees and the densities of large live trees) as the explanatory variables.

## Results

The dominant species in the stand in terms of both basal area and density was *T. diversifolia* (Table 1). Three conifers (*T. diversifolia*, *A. mariesii*, and *A. veitchii*) contributed 99.4% of the basal area and 86.0% of the density. The skew of the DBH distribution of *T. diversifolia* and *A. mariesii* was similar, while that of *A. veitchii* was much higher. This difference indicates that the DBH distribution of *A. veitchii* was considerably more biased toward smaller size classes than the distributions of the other two species. The DBH distribution differed significantly among the three species (Kruskal-Wallis test:  $P < 0.01$ , Bonferroni's multiple comparison:  $P < 0.01$  for each combination). Unlike the other two species, the basal area value for *A.*

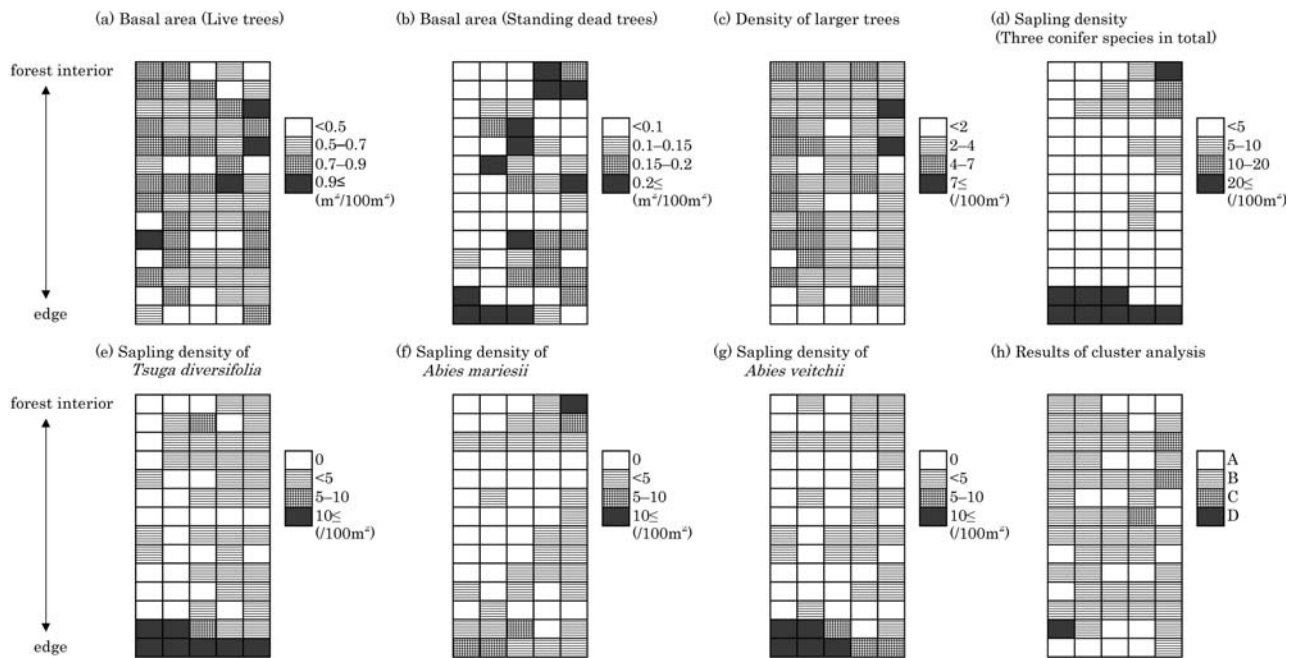


FIGURE 1. Dispersion maps of attributes of stand structure and the results of cluster analysis in the study stand. The cluster analysis was performed using a data set that included the basal areas (BAs) of both live trees and standing dead trees, and the densities of both larger live trees (DBH > 31cm) and saplings of a total of 3 species. The dendrogram using the average linkage between groups was separated into 4 groups at 15% of the squared Euclidean distance (similarity). The 4 groups were randomly assigned the letters A, B, C, and D.

*veitchii* was considerably lower than that of density. In each of the three species, the DBH of live trees was significantly smaller than that of the standing dead trees (Mann-Whitney U-test; each  $P < 0.01$ ).

Based on attributes of stand structure and regeneration, the edge effect varied with distance from the road. The basal area of standing dead trees was larger and the density of saplings was greater near the road (Fig. 1). Saplings of both *T. diversifolia* and *A. veitchii* were abundant near the edge, while those of *A. mariesii* were not (Fig. 1[e], [f], [g]). Cluster analysis showed that quadrats in the interior of the stand were structurally similar to those near the edge (Fig. 1[h]). These quadrats, which belonged to group A, characterized by a high basal area of standing dead trees, few large trees, and high sapling density, were in the gap phase of forest development (Fig. 1).

Sapling density was negatively correlated to both the basal area of live trees and the density of large trees, and positively correlated to the basal area of standing dead trees (Table 2). Since the simple correlation coefficients between the density of large conspecific trees and sapling density were not significant for each combination, sapling density was not dependent on the density of conspecific trees (Table 3). The results of multiple regression analysis with sapling density as the dependent variable showed that the factors affecting regeneration differed by species (Table 4). The sapling density of *T. diversifolia* and *A. veitchii* was strongly affected by the basal area of standing dead trees. Both the standardized partial correlation coefficient and the coefficient of determination ( $R^2$ ) of *A. veitchii* were higher than those of *T. diversifolia*. This model of multiple regression analysis was not significant for *A. mariesii*, indicating that the stand characteristics of saplings of this species differ from those of the other species.

## Discussion

### FOREST DEVELOPMENT NEAR EDGES

Canopy openings resulting from the death of canopy trees caused by natural (e.g., Taylor and Qin, 1992; Kubota, 1995) and human

(e.g., Kneeshaw et al., 1998; Nagaike et al., 1999; Van Pelt and Franklin, 1999) disturbances change stand structure and improve the light environment in the understory. The positive correlation between sapling density and the basal area of the standing dead trees that resulted from road construction implies that the death of canopy trees improved the light environment under the canopy and created gaps (e.g., Nakashizuka, 1985; Canham et al., 1990, Tables 2 and 4). It is widely recognized that some shade-tolerant conifers undergo advanced regeneration (i.e., form seedling and sapling banks; e.g., Greene et al., 1999; Messier et al., 1999). I found no correlation between the densities of conspecific trees and saplings, suggesting that regeneration depends on improvement of the light environment following the formation of canopy gaps rather than on advanced regeneration (Table 3).

As shown in Figure 1, it is difficult to quantify the extent of the edge effects from the road because they depend on attributes of stand structure. Cluster A in Figure 1(h) indicates the gap phase in the course of forest development; this is the initial phase of regeneration and is characterized by recruitment (Watt, 1947; Nakashizuka, 1984). As Murcia (1995), Cadenasso et al. (1997), and Esseen and Renhorn (1998) have pointed out, edge effects may be caused by interactions

TABLE 2  
Simple correlation coefficients among stand attributes

|                                     | Basal area<br>(live trees) | Basal area<br>(standing dead trees) | Density of<br>larger trees |
|-------------------------------------|----------------------------|-------------------------------------|----------------------------|
| Basal area<br>(standing dead trees) | -0.554**                   |                                     |                            |
| Density of larger trees             | 0.732**                    | -0.506**                            |                            |
| Sapling density                     | -0.279*                    | 0.484**                             | -0.360**                   |

Significant levels: \*\*  $P < 0.01$ , \*  $P < 0.05$ .

TABLE 3

Simple correlation coefficients between the densities of larger trees and saplings in the stand

|                           | Density of larger trees   |                       |                       |
|---------------------------|---------------------------|-----------------------|-----------------------|
|                           | <i>Tsuga diversifolia</i> | <i>Abies mariesii</i> | <i>Abies veitchii</i> |
| Sapling density           |                           |                       |                       |
| <i>Tsuga diversifolia</i> | −0.119                    | 0.102                 | —                     |
| <i>Abies mariesii</i>     | 0.277                     | −0.260                | —                     |
| <i>Abies veitchii</i>     | −0.057                    | −0.049                | —                     |

No combinations were significant ( $P > 0.05$ ).

The density of larger *Abies veitchii* could not be analyzed because it was constant in each grid.

among several variables. Therefore, it is necessary to evaluate edge effects using data sets that combine several variables, as Fraver (1994) and this study did.

#### DIFFERENCES IN SPECIES' RESPONSES FOLLOWING THE DEATH OF CANOPY TREES

The DBH distribution of *A. veitchii* was more skewed and had a smaller average than the distributions of *T. diversifolia* and *A. mariesii* (Table 1). Therefore, smaller size classes were more abundant in *A. veitchii* than in the other two species, showing that *A. veitchii* was regenerating successfully. Regeneration of *A. veitchii* continued between 1999 and 2001 (Nagaike, Abe, and Arai, unpublished data). The distribution of sapling density of each species differed, indicating that response to the improvement in light conditions varied by species (Fig. 1). The diameter and apical growth of *A. veitchii* exceed those of *T. diversifolia* in gaps (Nakamura and Obata, 1982; Kanzaki, 1984).

Kanzaki (1984) described the regeneration processes in a *T. diversifolia*–*A. veitchii* forest as follows: (1) after a disturbance, *Abies* regenerates from seedling and sapling pools and grows rapidly; subsequently, (2) pure stands of *Tsuga* remain after the death of *Abies*, which has a shorter life span. Yamamoto (1993, 1995) showed that *A. mariesii* and *A. veitchii* frequently recruited after gap formation, whereas *T. diversifolia* did so less frequently. Nakamura and Obata (1982, 1985) and Nakamura (1992) also pointed out that *A. veitchii* could form patches of saplings beneath canopy gaps in a *T. diversifolia* forest due to the difference in the survival ratio of seedlings and their positive response to increased light. In a *T. diversifolia*–*A. veitchii* forest near this study stand, regeneration was dominated by *A. veitchii* (Maeda et al., 1998). In *Abies*–*Tsuga*–*Chamaecyparis* forests of British Columbia, *A. amabilis* gap-fillers (trees that replace dead individuals in gaps) were more abundant than *T. heterophylla* or *T. mertensiana* (Lertzman, 1995). Therefore, it seems that *Abies* spp. generally outperform *Tsuga* during the sapling phase, suggesting that their growth response to increased light is better (Table 1).

Kohyama (1984) and Kimura et al. (1986) reported that *A. veitchii* saplings grew taller more quickly than *A. mariesii* and that the rapidly growing *A. veitchii* dominated *A. mariesii* in simultaneously decayed stands. *Abies mariesii* is more shade tolerant than *A. veitchii* (Kimura, 1963), and the photosynthetic response to improvement of the light environment at the leaf level is better in *A. veitchii* than in *A. mariesii* (T. Tanaka, unpublished data). The sapling density of *A. veitchii* in the study stand exceeded that of *A. mariesii* (Table 1) and was strongly affected by the basal area of standing dead trees (Table 4). Moreover, the model of multiple regression analysis in the case of *A. veitchii* was the most suitable, suggesting that saplings of this species were more responsive to the death of canopy trees following road construction.

It may be difficult to show an edge effect in the species composition of saplings. The death of canopy trees following road

TABLE 4

Results of multiple regression analysis with sapling density as the dependent variable

|                                  | Standardized partial correlation coefficient |                       |                       |             |
|----------------------------------|----------------------------------------------|-----------------------|-----------------------|-------------|
|                                  | <i>Tsuga diversifolia</i>                    | <i>Abies mariesii</i> | <i>Abies veitchii</i> | All species |
| Basal area (live trees)          | 0.257                                        | 0.060                 | 0.255                 | 0.144       |
| Basal area (standing dead trees) | 0.379*                                       | 0.341                 | 0.634**               | 0.443**     |
| Density of larger trees          | −0.332                                       | 0.050                 | −0.317                | −0.239      |
| F                                | 3.045                                        | 0.959                 | 8.774                 | 7.487       |
| R <sup>2</sup>                   | 0.198                                        | 0.087                 | 0.451                 | 0.260       |
| Significant                      | *                                            |                       | **                    | **          |

Significant levels: \*\*  $P < 0.01$ , \*  $P < 0.05$ .

construction alters the light environment in the understory, and *A. veitchii*, which is more shade intolerant than the other two species and quickly responds to increased light, may become the dominant species. While the density of canopy *A. mariesii* was greater than that of *A. veitchii*, the sapling density was lower. These results, which are likely caused by road construction, differ from those of a study conducted by Kohyama (1984) in a primary forest. He demonstrated that *A. mariesii* dominates *A. veitchii* in the understory of any mature stand and that *A. veitchii* is the competitively superior species in seedling-bank regeneration, while *A. mariesii* is superior in sapling-bank regeneration. Although the study stand is currently a mixed *T. diversifolia* forest, successful regeneration of *A. veitchii* may eventually change the species composition until a successor forms a canopy over *A. veitchii*. These results suggest that road construction affects not only stand structure but also species composition.

In conclusion, even though the road was constructed 30 yr before the time of this study, the stand structure near the edge still resembles that of gaps, and the species composition in the stand may have been changed by the road construction. Kohyama and Fujita (1981) and Kohyama (1982) showed in an *Abies* forest in central Japan that patches with a high density of saplings, which seem to indicate a gap phase, advance to a self-thinning stage. After this stage, it is unknown whether the original species composition of the stand is restored. The community dynamics (changes in species composition and density) in the study stand should therefore be monitored.

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